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*Identification of clonemates and genetic lineages using next-generation sequencing (ddRADseq) guides conservation of a rare species, *Bossiaea vombata* (Fabaceae)*

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1 **Identification of clonemates and genetic lineages using next-generation sequencing**
2 **(ddRADseq) guides conservation of a rare species, *Bossiaea vombata* (Fabaceae).**

3

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8

9 **Highlights**

- 10 • *Bossiaea vombata* was found to consist of five clones using ddRADseq. Three sites were
11 monoclonal, and two clones were identified at the fourth and largest site.
- 12 • The application of genetic thresholds allowed non-identical but similar multilocus
13 genotypes to be assigned as clonemates reducing the risk of over-estimating clonal
14 diversity.
- 15 • ddRADseq was an efficient, cost-effective method for clonal identification in a non-model
16 plant and applicable to phylogenetic studies of a group of leafless bossiaeas from eastern
17 Australia using a single library.
- 18 • ddRAD loci from the same library may have limited utility for chloroplast haplotyping
19 due to attributes inherent in library preparation and/or the absence of a reference genome
20 from *B. vombata* or a closely related species.

21

22

23 **ABSTRACT**

24 Plant species capable of clonal reproduction range from rare, sterile species that have a high
25 extinction risk to invasive plants that influence the structure of ecosystems. There is increasing
26 evidence that clonality in combination with reduced fecundity and limited dispersal capacity

27 increases extinction risk. As many conservation targets are not well-characterised genetically, our
28 objective was to determine the utility of sequencing a reduced representation of the genome to
29 inform the conservation of a non-model plant species. We sequenced a single DNA library from
30 a recently described, rare, clonal species; *Bossiaea vombata*. Multiple assemblies of that library were
31 used to evaluate our ability to assess genetic variation and clonal assignment, to identify
32 chloroplast haplotypes and to obtain phylogenetic information. Next generation sequencing
33 (ddRADseq) provided a cost and time effective method for identifying clones and assigning
34 clonemates despite increased levels of missing data in comparison to more traditional methods
35 (e.g. microsatellites). We applied a threshold of genetic difference to determine whether
36 individual samples belonged to the same or different clones. *Bossiaea vombata*, was found to
37 comprise only five clones with all but one site being monoclonal - indicating that the genotypic
38 diversity of the species cannot be determined from a census of stems. The ddRADseq method
39 showed utility for phylogenetic analysis but we identified possible shortcomings in chloroplast
40 haplotyping using loci from the same library.

41 **Keywords:** chloroplast haplotype; genet; multilocus lineage; low fecundity; next generation
42 sequencing; population structure

43

44 1. Introduction

45 The field of conservation biology has focused predominantly on sexually reproducing species
46 although conservation requirements can differ between clonal and sexual species. On paper,
47 clonal species with limited fecundity are poor candidates for conservation but we know that
48 clones can survive many times the generation span of their sexual conspecifics and related taxa
49 (e.g., Lynch and Balmer, 2004; Rossetto et al., 1999; Mock et al., 2012). Population fitness and
50 adaptability are dependent on genetic diversity which is a proxy for species health (Mable, 2019).
51 The extended longevity of a genet increases the chance of occasional sexual recruitment and can
52 moderate genetic decline because the loss of individual ramets does not mean the loss of a genet
53 (Jongejans et al., 2008; de Witte and Stöcklin, 2010). However, if asexual reproduction reduces
54 opportunities for genetic novelty via meiotic recombination, subsequent adaptive capacity may
55 also be reduced (Hoffmann and Sgrò, 2011).

56 A changing environment is likely to disproportionately affect species that are constrained either
57 genetically or geographically (Fordham et al., 2013). The resulting abundance of many species
58 will rest on their capacity to respond either by adapting to new conditions in their current
59 environments or avoiding them by dispersing to new areas (Hoffmann and Sgrò, 2011). Cunze et
60 al., (2013) predict that most plant species will decline when faced with contemporary climate
61 change as dispersal limitations restrict their capacity to shift geographic ranges. If true, species
62 reliant on clonal growth, including *Bossiaea vombata* Ross, 2018, the focal species of this study,
63 form a group with an elevated extinction risk because their rate of population expansion and/or
64 ability to shift range due to limited dispersal is generally more restricted compared to their
65 sexually reproducing counterparts. However, the success and persistence of clonal plants over
66 historic periods of climate change suggest alternative adaptive strategies are also in play. One
67 such strategy suggested by Dodd and Douhovnikoff (2016), is the accumulation of epigenetic
68 changes, as an alternative to DNA mutation and recombination that may imbue clonal plants
69 with a rapid response to environmental variation by enabling phenotypic plasticity and selection
70 in situ.

71 As a pre-requisite to the management of clonal plants for conservation, their capacity to disperse,
72 adapt and balance sexual and asexual reproduction must be addressed. Ideally, these
73 conservation management strategies would incorporate evolutionary principles (Sgrò et al., 2011;
74 Christmas et al., 2015), highlighting the need to expand genetic studies of clonal species. The loss
75 of reproductive capacity is considered a threatening process for species survival and the extent
76 of clonality within a species can be a factor in its fecundity (Eckert, 2002; Silvertown, 2008).
77 Knowledge of the underlying patterns of genetic variation and the genetic lineages generated by
78 historic population processes is of particular relevance when planning conservation strategies
79 (Young et al., 1996; Hoffmann and Sgrò, 2011), particularly for species where clonality can have
80 a negative impact on the production of new genotypes (Charpentier, 2002).

81 Within 50 years most Australian ecosystems are expected to experience climatic changes resulting
82 in environmental stress for organisms that are adapted to current conditions (Dunlop et al., 2012).
83 The extent and distribution of clonality among most of Australia's 22,000+ vascular plants is
84 unknown; however, data is accumulating on its association with rare, relict species with reduced
85 fecundity or dysfunctional breeding systems (for example, Coates, 1988; Lynch et al., 1998, Sydes
86 and Peakall, 1998; Bartolome et al., 2002; Kimpton et al., 2002; Peakall et al., 2003; Gross and
87 Caddy, 2006; Gross et al., 2011). The assumption that clonal growth enables local persistence
88 when sexual reproduction is suppressed has been borne out by empirical studies but data are
89 scarce for Australian species (e.g. Rossetto et al., 1997; Coates et al., 2002; James and McDougall,
90 2007; Millar et al., 2010).

91 Genetic studies of species capable of clonal reproduction become important if, for example, a
92 demographic census of populations is likely to misrepresent the effective population size
93 (Arnaud-Haond et al., 2007). The expanding use of genomic techniques has shifted the focus of
94 population studies from a small number of neutral genetic markers or candidate genes to
95 genome-wide screening of genetic variation of evolutionary, adaptive and conservation
96 importance (Ouborg et al., 2010; McMahon et al., 2014). The use of single nucleotide
97 polymorphisms (SNPs) is highlighting the evolutionary continuum that links lineages to
98 populations and species (Coates et al., 2018) but clonal plants have not been a target of these

99 studies to date. Genomic data can improve our phylogenetic understanding of species emergence
100 and provide an evolutionary context for conservation efforts (Byrne, 2007). McMahon et al., (2014)
101 propose the development of 'conservation priors' based on genomic data to guide management
102 of populations, and decision frameworks for biodiversity conservation are also being developed
103 that utilise genomic information in an evolutionary context (Allendorf et al., 2010; Hoffmann et
104 al., 2015).

105 Deciding on an appropriate method for the generation of genome-wide data for clonal species
106 identified as conservation candidates can be influenced by several factors. The availability of
107 existing data (e.g. reference genomes or transcriptomes or microsatellite loci) reduces the time
108 taken to identify informative genetic regions but is limited or unavailable for most species.
109 Microsatellites have traditionally been the marker of choice for genetic studies of clonal plants
110 but their utility for taxa where genetic diversity is low, is determined by whether an adequate
111 number of polymorphic loci can be identified, and cross-species amplification may not be
112 adequate for comparative studies. When few or no data exist for a species, the most cost effective
113 and productive approach may be to sequence a reduced representation of the genome to identify
114 SNPs (Elshire et al., 2011; Peterson et al., 2012). Reduced representation sequencing has become
115 increasingly accessible for those non-model species that have become conservation targets
116 (Andrews et al., 2016). Although Pantoja et al., (2017) used SNPs to characterise genotypic
117 variation and population structure in a facultative clonal herb, there have been few studies
118 reporting the use of SNPs in the assessment of clonality.

119 Here, we evaluate the utility of genome-wide SNP data obtained via double digest restriction site-
120 associated DNA sequencing (ddRADseq, Peterson et al., 2012) to address questions relevant to
121 the conservation of a non-model, clonal plant. *Bossiaea vombata* was selected for analysis due to
122 its endangered conservation status. The species has been confirmed to occur within an area of
123 less than 5 km² in the Wombat State Forest, Victoria, Australia. Our primary aim was to
124 investigate the number of clonal individuals that the species contains - thus, the extent of clonality
125 among all known discrete patches and, therefore, the vulnerability of *B. vombata* to extinction. We

126 also aimed to assess the utility of the same dataset to address phylogenetic- and cpDNA haplotype-
127 based questions relevant to conservation.

128

129 2. Materials and Methods

130 *Bossiaea vombata* (Fabaceae) is one of ~78 species in the endemic Australian genus, *Bossiaea*. Most
131 species are small, woody shrubs to about 1 – 2 m. In one group, the “leafless” bossiaeas, leaves
132 are reduced to scales. Generally, this life form is considered to be an adaptation to harsh dry
133 conditions in arid areas in Australia but most leafless bossiaeas are found in higher rainfall areas
134 with annual precipitation >500 mm. Thompson (2012) recognised 12 eastern Australian leafless
135 *Bossiaea* species and assigned them to one of four subgroups. Of the 12 species, six were described
136 between 1808 and 1879 and six species were described relatively recently - between 2008 and
137 2012. Five of the post-revision species were described from *B. bracteosa* (Ross, 2008; McDougall,
138 2009 and one from *B. ensata* (Thompson, 2012, and key therein). Clonal growth has been noted in
139 ten species and can be extensive (Thompson, 2012). The recently recognised species have very
140 restricted geographic ranges which makes the whole group a candidate for conservation.

141 *Bossiaea vombata* is known from four sites comprising a total combined patch area of <500 m²
142 spread over a maximum distance of 3 km. The species reproduces rhizomatously and flowers in
143 spring (Fig. S2, S3). The longevity of individuals is unknown but at least one site presumed to
144 contain ramets of a single plant has been known prior to 2008 (Ross, 2008). Only site 4 contained
145 more than one patch and stems were less dense in patches 4 and 5 compared to other patches
146 (Fig. S4). The breeding system is unknown. Pollen viability ranged from 15 – 90% in spring 2018
147 (E. James, unpublished data) suggesting that for at least some sites, pollen viability should be
148 sufficient for sexual reproduction in the absence of other impediments such as self-
149 incompatibility. Anecdotal reports of the presence of pods with undeveloped seed were
150 confirmed during this study (personal observations, pers. comm Donna Thomas, Ballarat Botanic
151 Gardens, Victoria) but seedlings have never been observed so the species is considered
152 functionally sterile. The listing of *B. vombata* as Endangered under Victorian State Government
153 Legislation provides some protection and is appropriate given the extremely restricted confirmed

154 distribution and very small combined patch size. The species is highly vulnerable to threats such
155 as fire and damage from vehicles and road maintenance activities due to the small number (and
156 size) of patches, their occurrence on road verges and the apparent lack of seedling recruitment.

157 2.1 Sampling

158 As *B. vombata* is leafless, 40 samples of young stem tissue were collected for genetic analysis across
159 all known locations in the Wombat State Forest, Victoria, Australia (under DELWP permit
160 10007717; Table 1, Fig. 1). Five samples were collected from each of eight discrete patches located
161 across four sites. Because of the clonal nature of the species, we refer to four “sites” to indicate
162 general location and describe discrete groups of stems with a minimum distance between them
163 of 7 m as “patches”. We estimated the size of each patch with geolocations recorded at their
164 centres. Samples were taken from four equidistant points around the perimeter and one central
165 point of each patch, wrapped in labelled aluminium foil and kept on ice until they were snap
166 frozen in liquid N₂ within 5 h and stored at -20 °C. All sites consisted of a single patch except Site
167 4 where five patches occurred within an area of ca.1000 m². In addition, three samples of a single
168 *B. riparia* individual and one each of *B. bracteosa*, *B. grayi* and *B. walkeri* were included for
169 comparison (cultivated at RGB Victoria; Table 1).

170 [Fig. 1]

171 [Table 1]

172 2.2 DNA isolation

173 Frozen stem material (~100 mg) was ground to a fine powder using a mortar and pestle with
174 liquid N₂ and the addition of a small amount of sand. Genomic DNA was isolated using the
175 manufacturer’s CTAB protocol for Bioline ISOLATE II Plant DNA Kit except that DNA was
176 **eluted in a total volume of 40 µL**. DNA quality was confirmed by using 1.5 % agarose gel
177 electrophoresis in 1xTBE buffer for 45 min at 100 V and stained with SybrSafe (Invitrogen,
178 Thermo Fisher Scientific Australia). DNA isolations were quantified using a Qubit v3.0
179 fluorometer (Invitrogen Life Technologies) and stored at -20 °C.

180 2.3 Library preparation and sequencing

181 A modified version of the Peterson et al., (2012) ddRADseq protocol was used to prepare DNA
182 libraries. A detailed protocol for preparation and amplification is available at
183 michaelamor.com/protocols. Our final DNA library contained 53 samples including nine
184 technical replicate pairs. In summary, genomic DNA was digested for 18 hours with EcoRI-HF
185 and AseI restriction enzymes (New England Biolabs), barcoded adapters were ligated to digested
186 DNA fragments. Non-ligated adapters were removed before libraries were size-selected by
187 magnetic bead purification using Jetseq Clean (Bioline)/PEG 8000 buffer solution at 0.5x then 0.9x
188 the DNA solution volume.

189 PCR-based indexing of the individual libraries was conducted using real-time PCR (rtPCR) in a
190 Bio-Rad CFX96 thermocycler, with amplification stopped after 10 cycles in total. Amplification
191 adequate for sequencing, whilst minimising PCR bias, was assessed by visualising fluorescence
192 curves and ensuring that they had not plateaued. Amplified libraries were pooled in equal
193 concentrations based on Relative Fluorescence Unit outputs from rtPCR and
194 concentrated/purified using Jetseq Clean beads/PEG 8000 buffer solution (1.8x DNA solution
195 volume). The pooled library was size-selected at 350-550 bp in a Pippin Prep (Sage Science) using
196 a 2% agarose (100-600 bp) cassette and quantified via qPCR using a Jetseq Library Quantification
197 Hi-ROX kit (Bioline) on a Bio-Rad CFX96 thermocycler. The library was diluted to 10 nM,
198 denatured and diluted to 20 pM for sequencing using a 600 cycle (paired-end) v3 MiSeq Reagent
199 Kit on an Illumina MiSeq with 10% PhiX spiked into the run.

200 2.4 Quality filtering and bioinformatics pipeline

201 Raw paired-end reads were trimmed if the quality dropped below a score of phred20, based on
202 a sliding-window of four bases using Trimmomatic v0.38 (Bolger et al., 2014). Reads below our
203 minimum length requirement of 250 bases were discarded. Finally, reads were trimmed if
204 Illumina adapters were present, filtered for microbial and fungal contaminants using Kraken
205 v2.0.6 (Wood and Salzberg, 2014) and merged using PEAR v0.9.8 (Zhang et al., 2014). Merged
206 and unmerged reads (read one only) were demultiplexed into individual sample read-sets using
207 the 'process_radtags' feature of STACKS v1.4.6 (Catchen et al., 2013).

208 2.5 De novo assembly of RAD loci

209 We performed *de novo* assembly of unclassified RAD loci using ipyrad v0.7.29 (Eaton, 2014).
210 Further sequence quality filtering was performed to convert base calls with a score of <30 into Ns,
211 whilst excluding reads with ≥ 15 Ns. A minimum depth of six reads per individual was required
212 for clustering of putative loci. Loci containing more than two alleles were excluded as potential
213 paralogs.

214 Three datasets were assembled to address independent questions. 1) clonal identification
215 assembly (n=37): All *B. vombata* individuals with suitable read numbers, $\geq 25/37$ individuals
216 present at each locus; 2) chloroplast mapping assembly (n=13): 2-3 *B. vombata* individuals per site,
217 $\geq 5/13$ individuals present to retain each locus; 3) phylogenetic assembly (n=12): one *B. vombata*
218 individual per patch (n=8) plus one individual each from *B. bracteosa*, *B. grayi*, *B. riparia* and *B.*
219 *walkeri*, retaining only loci with $\geq 6/12$ samples present.

220 For assembly 1, reads were clustered at 99% similarity within and among each individual (2.5
221 [250 bp locus] to 5.6 [560 bp locus] SNPs per locus). **This similarity was selected for several**
222 **reasons; (i) reducing clustering similarity to 95% (allowing for up to 28 SNPs per locus) did not**
223 **substantially alter the number of loci assembled, suggesting that all assembled loci had**
224 **relatively few SNPs, (ii) we only required a single SNP per locus for analyses, (iii) this level**
225 **of similarity encompassed variation between technical replicate pairs and, finally, although**
226 **population/species-level studies typically use comparatively more relaxed similarity**
227 **clustering, (iv) we considered 2-5 SNPs per locus as a suitable number as we aimed to**
228 **investigate finer-scale diversity among clones/individuals. For assemblies 2 and 3, we**
229 **clustered reads at a similarity level of 95% to allow for diversity among species.**

230 To remove potentially linked SNPs in all assemblies, we used a '--thin 1000' filter in VCFtools
231 v0.1.15 (Danecek et al., 2011). This allowed us to retain only the first SNP when multiple SNPs
232 occurred within 1,000 bases of each other, which was greater than our largest locus length. The
233 resulting VCF files were converted to 'genind' objects using the 'vcfR2genind' command. Finally,
234 we assessed the occurrence of outlier loci using Bayescan v2.1 (Foll and Gaggiotti, 2008) and a
235 PCA approach implemented via the 'pcadapt' package (Luu et al., 2019) in R.

236 2.6 Assigning clones

237 Ramets from a single genet may differ in their genotypes due to genotyping errors and somatic
238 mutation and closely related but different clones may exhibit very similar genotypes. The
239 application of thresholds is typically applied to clonal data to determine how much variation is
240 acceptable between clonemates (Douhovnikoff and Dodd, 2003; Obrien et al., 2014).

241 We used multiple analyses to investigate potential thresholds among multi-locus lineages and
242 identify *B. vombata* clones. First, we calculated a pairwise genetic distance matrix among *B.*
243 *vombata* individuals using the 'bitwise.dist' function in the 'poppr' package (Kamvar et al., 2014)
244 in R. A histogram and scatter plot of the genetic distance matrix was used to identify potential
245 threshold values for *B. vombata* to distinguish clonemates from distinct multilocus lineages
246 (Meirmans and Van Tienderen, 2004). We constructed an ultrametric tree ('upgma' function in
247 'phangorn' package; Schliep, 2011) based on the same distance matrix and overlaid a
248 discrimination threshold based on the previous graphs (Kamvar et al., 2015). We used this
249 combined approach to determine the number of clades supported at the point of our determined
250 threshold. Finally, we investigated the number of genetic clusters, via Discriminant Analysis of
251 Principal Components (DAPC; Jombart et al., 2010) using the 'adegen' package (Jombart, 2008),
252 without a priori site input. The number of genetic clusters supported via DAPC was determined
253 via the Bayesian Information Criterion (BIC; Schwarz, 1978).

254 2.7 Genetic diversity and demographics

255 To investigate clonal diversity we calculated F_{IS} and pairwise F_{ST} via the 'boot.ppfis' and
256 'pairwise.fst' functions, respectively, in the 'hierfstat' R package (Goudet 2005). Pairwise G_{ST} was
257 calculated via the 'pairwise_Gst_Hedrick' function in the R package 'mmod' (Winter 2012).
258 Global observed heterozygosity (H_o) and gene diversity (expected heterozygosity, H_e) were
259 calculated using the 'basic.stats' function in 'heirfstat'.

260 2.8 Chloroplast haplotypes from ddRADseq data

261 We tested the viability of identifying chloroplast haplotypes using ddRADseq data in place of
262 amplicon sequencing using assembly 2. We mapped reads from 2-3 individuals per identified
263 clone (those with the highest read count) with ipyrad to a consensus sequence generated from
264 the alignment of two chloroplast genomes from the family Fabaceae: *Glycine max* (GenBank

265 accession: CM010429.1) and *Indigofera tinctoria* (GenBank accession: KJ468098.1). Geneious v9.1.8
266 (<https://www.geneious.com>) was used to perform the alignments and to visualise the resulting
267 phylip sequence files.

268 2.9 Suitability of ddRADseq for phylogenetic study of eastern Australian leafless bossiaeas
269 We used ipyrad to construct the phylogenetic assembly *de novo*. Similarity thresholds for within
270 and among sample assembly was set to 95%. All other parameters were defined as above. The
271 resulting phylip file was used to produce a Maximum Likelihood phylogeny via RAxML v8.2.9
272 (Stamatakis, 2014) based on the GTR+G model of evolution with 1,000 bootstrap replicates.

273

274 3. Results

275 3.1 Sequencing run statistics

276 Sequencing of our ddRADseq library resulted in 8.7 million paired-end reads that contained a
277 maximum of a single error per 1,000 bases (Phred quality score of 30). Of the quality paired-end
278 reads, approximately 7.8 million (90%) were retained after trimming for quality and Illumina
279 adapters. Less than 1% of these reads were filtered out by attempting to classify them to microbial
280 and fungal databases. On average, 30% of the retained reads overlapped sufficiently and were
281 combined into 'single end data'. We retained only read one of the non-merged data (Unmerged
282 read two sequences were discarded as they were considered linked to their associated read one
283 sequence and were lesser quality). Merged and non-merged (read one only) reads (5.1 million
284 total) were demultiplexed according to their barcodes, which resulted in 4.96 million reads being
285 retained (97%). The mean number of reads per sample was 93,505 (SD=24,249) after discarding
286 three samples with insufficient read numbers (<1%, relative to read numbers of other samples)
287 (see table 1) that were attributed to error during library preparation process.

288 3.2 Assembly statistics

289 The minimum read depth (number of identical reads) per individual of each dataset was 6. Across
290 all three assemblies, an average of 2% of our total putative loci were filtered out as they did not
291 meet our set quality criteria. An average of 91% of loci were filtered out as they did not occur

292 within our threshold for minimum sample number. Approximately 1% of our final loci were
293 filtered from all three assemblies due to treating loci with >2 alleles as potential paralogs.

294 Across all *B. vombata* individuals used for clonal identification (assembly 1), the average read
295 depth per site was 8.3. This assembly produced 1,368 loci with 755 informative sites. Further
296 filtering gave us 720 unlinked SNPs with 720 informative sites. The average read depth for our
297 chloroplast dataset (assembly 2) was 9.2. The assembly produced 8 loci with 3 informative sites.
298 In our phylogenetic dataset (assembly 3), the average read depth per site was 9.6. This assembly
299 resulted in 2,893 loci with 2,866 informative sites (minor alleles occurred in >1 sample). *Bossiaea*
300 *vombata*, *B. bracteosa*, *B. grayi*, *B. riparia* and *B. walkeri* individuals were represented in 63-93%,
301 27%, 11%, 29% and 13% of assembled loci, respectively. BayeScan identified zero outliers and 28
302 outliers were detected by pcadapt. As no outliers were common across both approaches, we
303 considered those identified by pcadapt to be false positives and all following analyses were
304 conducted using our three complete datasets.

305 3.3 Clonal assignment

306 We used a genetic distance-based approach to visualise a percentage threshold between diversity
307 levels that reflect i) library preparation/sequencing error and somatic mutations within a clone
308 and ii) inter-clonal variation representing multi-locus lineages. A clear valley was present at 0.4%
309 genetic divergence (Fig. 2a). This threshold value was overlaid on our ultrametric topology and
310 allowed us to identify five lineages which we treated as putative clones (Fig. 2b).

311 [Fig. 2]

312 The distribution of clones of *B. vombata* revealed that sites 1, 2 and 3 were monoclonal and each
313 consisted of a single patch identified as a unique genet ranging in area from ca. 8 – 100 m². **The**
314 **largest distance between clones was approximately 3 km (between sites 1 and 4).** The fourth
315 site, covering the greatest area, contained two clones. **One large fragmented clone dominated**
316 **the site with ramets identified at least 40 m apart.** Two samples situated along a trackside edge
317 of patch four and adjacent farmland belonged to clone five **with ramets a minimum distance of**
318 **2 m from the dominant clone.** There was no correlation between the geographic distance between
319 sites and the genetic similarity between genets including the two clones from site 4.

320 DAPC analysis also supported the presence of five *B. vombata* clones in the Wombat State Forest
321 (Fig. 3a). The probabilities of assignment to a particular group (clone) for each individual were
322 all 100% (Fig. 3b). These results were congruent with those obtained using genetic thresholds.

323 [Fig. 3]

324 3.4 Population structure and diversity

325 The population structure observed via DAPC did not reflect the geographical distribution of our
326 clones within the Wombat State Forest. **This may reflect a lack of modern admixture, historical**
327 **loss of genetically intermediate individuals or may be an artefact of few viable**
328 **populations/sites**. Pairwise G_{ST} values supported this finding as differences reported among
329 clones did not correspond to proximity. Population differentiation values ranged from 0.2233 to
330 0.5583 and supported distinction among five clones (Table 2).

331 [Table 2]

332 Each clone showed an excess of heterozygotes, as all F_{IS} values were negative (clone 1: -4.12315,
333 clone 2: -0.61055, clone 3: -0.8436, clone 4: -0.7785 and clone 5: -1.0749). We note that clones one
334 and five had the most extreme negative F_{IS} values and had the fewest samples ($n=2$ each). Genetic
335 variability of *B. vombata*, inferred by H_o and H_e , was relatively low; 0.220 and 0.130, respectively.

336 3.5 Chloroplast haplotypes

337 Mapping our chloroplast dataset (assembly 2) to our reference resulted in eight loci (2,738
338 nucleotides combined, <2% of the reference genome) all of which were conserved across *B.*
339 *vombata* samples and references. None of the loci corresponded to sequences for cpDNA regions
340 trnL-F and ndhF that exhibit variation among other eastern Australian leafless *Bossiaea* species
341 (available on GenBank).

342 3.6 Phylogenetic relationships

343 **Phylogenetic relationships among species were largely well supported**. In our unrooted tree,
344 *B. bracteosa* was supported as the closest relative of *B. vombata* (Fig. 4b). *Bossiaea vombata* was
345 supported as a distinct taxon with no phylogenetic structure supported within the clade.

346

347

348 **4. Discussion**

349 We found the next generation sequencing method, ddRADseq (Peterson et al., 2012), to be an
350 efficient, cost-effective method for quantifying genetic variation and identifying clones in *B.*
351 *vombata* - a highly restricted plant species lacking available genetic information. We prepared and
352 sequenced a single library and performed three independent assemblies to determine that *B.*
353 *vombata* comprises five clonally reproducing genetic individuals (genets) that range in size from
354 ca. 8 – 102 m². We also obtained **eight** completely conserved cpDNA loci from *B. vombata* and
355 distantly related members of the family Fabaceae. Finally, using data from *B. vombata* and all
356 available close relatives, we were able to construct a phylogenetic tree with high bootstrap
357 support - showing promise for the ability to resolve phylogenies and investigate clonality from a
358 single library.

359 The identification of five distinct genets of *B. vombata* from the four known sites (three
360 monoclonal) indicates that the genotypic diversity of the species is extremely low and cannot be
361 determined from an in-situ census of stems. With so few genetic individuals, no site is of greater
362 conservation value than another and all genets should be considered a single ESU. While
363 ddRADseq is known to produce datasets with random missing data, the inclusion of technical
364 replicates and the application of a genetic threshold allowed us to minimise the risk of
365 overestimating clonal diversity.

366 Our results support clonal growth as the primary mode of reproduction at all four *B. vombata*
367 sites, a finding that is consistent with a lack of direct evidence of seedling recruitment despite
368 apparently viable pollen and casual observations of pod formation. In the absence of sexual
369 reproduction, differences in clonal fitness are likely to result in loss of genets over time (Honnay
370 and Bossuyt, 2005). The population structure of mostly monoclonal sites points to *B. vombata*
371 being a relict species with extant populations maintained via vegetative rather than sexual
372 reproduction (Kruckeberg and Rabinowitz, 1985). Negative F_{IS} is one signature of asexual
373 reproduction (Halkett et al. 2005). Stoeckel et al. (2006) showed that significant negative F_{IS} in

374 *Prunus avium* could be explained by asexual reproduction and Balloux et al. (2003) predict highly
375 negative F_{IS} values only for levels of clonality approaching 1, as found for three of the four *B.*
376 *vombata* sites. As F_{IS} estimates for all *B. vombata* clones were highly negative, our genetic results
377 support clonality as the primary mode of reproduction. The identification of a second, smaller
378 clone at the perimeter of one patch at the largest site, site 4, may be an indication of rare recent
379 recruitment, however, its position at the edge of a road and nearby cleared farmland may also
380 reflect the decline of most of this clone due to habitat loss.

381 Based on limited cpDNA data available (GenBank), variation is present among (and within some)
382 leafless *Bossiaea* species. Mapping ddRADseq reads to a chloroplast genome gave us the potential
383 to investigate cpDNA haplotypes without additional amplicon sequencing. As no chloroplast
384 genomes were available for closer relatives, we were unable to determine whether the large
385 phylogenetic distance between *B. vombata* and the reference sequence limited locus identification
386 to only highly conserved regions across the family and/or whether the ddRAD process also
387 contributed due to read length or choice of restriction enzymes. However, we consider that the
388 phylogenetic distance between our reference and the leafless bossiaeas combined with the
389 technical limitations of ddRADseq are likely to have contributed to our recovery of only
390 conserved regions.

391 We consider that mapping *Bossiaea* ddRADseq loci to chloroplast reference genomes from within
392 the leafless bossiaeas will be necessary before we can assess whether variable cpDNA loci are
393 recoverable and informative using the shorter reads from ddRADseq compared to amplicon
394 sequencing. A combination of careful enzyme selection, availability of an appropriate reference
395 sequence and longer read lengths produced from alternative sequencing platforms are
396 recommended for future ddRADseq experiments. The ddRAD process results in restriction
397 fragments that mostly originate from the nuclear genome due to the relative abundance of nuclear
398 compared to plastid restriction sites. However, the number of overall fragments obtained,
399 including from the chloroplast, is highly tunable. Therefore, the use of a single library for several
400 purposes, as in our study or specifically for haplotyping, may be limited for now, however, as

401 technology improves and cost-per-sequencing-read continues to decrease, the viability of this
402 approach may improve in the future.

403 Broadhurst et al., (2016) found that range size, life form and breeding system are some of the
404 important predictors of genetic diversity in Australian plant species and changing environmental
405 conditions are known to affect plant mating systems (Eckert et al., 2010). Reproductive strategies
406 influence the distribution of genetic diversity which has a strong effect on plant population fitness
407 and future viability (Christmas et al., 2015). Sexual failure in most exclusively clonal populations
408 is attributed to environmental factors (Vallejo-Marín et al., 2010), and sexual extinction can
409 eventuate where sexual reproduction is suppressed for a prolonged period and growth is almost
410 exclusively clonal. (Honnay and Bossuyt, 2005). If reproductive constraints lead to permanent
411 sterility within a species, its extinction is inevitable in the absence of asexual reproduction
412 (Silvertown, 2008). Loss of genotypes can lead to monoclonal populations of a species with
413 significant implications for population viability (Eckert, 2002; Vallejo-Marín and O'Brien, 2007).
414 This would appear to be applicable to *B. vombata*, adding to Australian examples of woody species
415 with functionally sterile monoclonal populations that include *Halagorodendron lucasii* (Sydes and
416 Peakall, 1998), *Acacia anomola* (Coates, 1988) *Santalum lanceolatum* (Warburton et al., 2000), *Lomatia*
417 *tasmanica* (Lynch et al., 1998) and *Grevillea renwickiana* (James and McDougall, 2014) all of which
418 are actively managed for conservation. However, analysis of the breeding system is required to
419 confirm sexual reproductive failure in *B. vombata* and identify its underlying basis. Investigating
420 barriers to sexual recruitment in-situ or using an ex-situ collection, using controlled pollinations
421 within and among clones, is feasible now that the location of clones is known.

422 The role of species with the capacity to buffer environmental fluctuations may be important in
423 stabilising systems as species suites change even if those individual species are uncommon and
424 rarely recruit from seed (Isbell et al., 2011). However, the current rarity of *B. vombata* limits its
425 possible ecological influence unless altered habitat conditions promote rapid clonal expansion or
426 sexual reproduction. The incremental loss of rare species will ultimately decrease unique
427 biodiversity in ecological systems, and in the absence of detailed ecological information, there is
428 potential for subsequent loss of species reliant on interactions with these vulnerable species for

429 survival. We have no direct observations relevant to the impact of climate change on *B. vombata*.
430 However, indirectly, the distribution of leafless bossiaeas in south-eastern Australia is consistent
431 with fragmentation influenced by historical climate processes including increasing aridity
432 (Quilty, 1994) and is likely to be impacted by ongoing changing climatic conditions (Dunlop et
433 al., 2012). Post-colonisation land use changes in the Wombat State Forest (gold mining, timber
434 harvesting, agriculture & grazing) and surrounding areas combined with changing temperature
435 and rainfall patterns in recent decades are likely to have had a significant impact on
436 environmental conditions in the areas where *B. vombata* is found now. In association with the
437 small number of clones and apparent functional sterility of *B. vombata*, those factors also have the
438 potential to affect its future persistence, without intervention, particularly if the dispersal capacity
439 of the species is limited to the rate of clonal growth. However, the small number of clones makes
440 it feasible to propagate each clone and develop an ex-situ population. This would enable study
441 of reproductive constraints to determine whether self-incompatibility or inbreeding is affecting
442 the production and viability seed given that three of the four sites are monoclonal, and the
443 remaining site has only two clones. If successful, artificial pollinations between clones may be an
444 effective way to increase genetic novelty and genotypic diversity via seed. Vegetatively
445 propagating multiples of each clone would allow the augmentation of extant populations with
446 clone/s already present at a site and raises the possibility of introducing a mix of all clones at new
447 locations to re-establish genetic connectivity.

448 The efficacy of ddRADseq data for phylogenetic studies is already well documented (Lemmon
449 and Lemmon, 2013; Ree and Hipp, 2015) and our results support its utility for phylogenetic and
450 clonal diversity studies within southeastern Australian leafless bossiaeas (and other clonally
451 reproducing species of broader interest; such as other plants, algae and corals). Developers of
452 biodiversity conservation policy acknowledge the need to retain dynamic evolutionary processes
453 (McGuigan and Sgrò, 2009; Sgrò et al., 2011) and biologically relevant conservation units can be
454 identified by characterising the evolutionary history of species and populations and clarifying
455 the genetic relationships between them. *Bossiaea vombata* is clearly a single ESU that should be
456 managed as such, but the relevant information is lacking for the remaining leafless *Bossiaea*. The
457 use of ddRADseq to generate genomic data from populations of all eastern leafless bossiaeas

458 would facilitate the identification of genetic lineages and the extent of clonality and clarify
459 phylogenetic relationships between currently recognised taxa. Such a genetic overview is one
460 step towards developing conservation guidelines for the southeastern Australian leafless
461 bossiaea group by taking into consideration the evolutionary processes that have shaped the
462 group to this point. Uncovering the relationships between lineages across all twelve leafless
463 bossiaeas in eastern Australia would provide a basis for identifying both ESUs (Moritz, 1994) and
464 MUs (Palsbøll et al., 2007), irrespective of current taxonomy, to design a conservation strategy
465 that recognises the evolutionary history of the whole group. Practical guidelines could include
466 hand-pollinations between natural populations or ex-situ plants to increase genetic connectivity
467 and as a seed source to establish new populations or to augment existing populations. Ultimately,
468 the decisions used to prioritise targets for conservation must account for their potential
469 evolutionary and ecological contribution, in addition to factors such as extinction probability, the
470 feasibility of removing threats and the efforts required to ensure their future.

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480 **Supplementary information**

481 The following is supplementary information for this article:

482 Supplementary information Amor et al.pdf

483 **CRedit author statement**

484 **Michael Amor:** Methodology, Software, Investigation, Validation, Formal analysis, Data
485 curation, Visualization, Writing- Reviewing and Editing. **Joshua Johnson:** Resources, Writing-
486 Reviewing and Editing. **Elizabeth James:** Conceptualization, Funding Acquisition, Project
487 Administration, Investigation, Visualization, Writing- Original draft preparation, Reviewing
488 and Editing, Supervision.

489 **Data accessibility**

490 Raw data and all working phylip/vcf files (post-filtering and assembly) are available from
491 GenBank (Accession: XXXXXXXXXXXX).

492

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664

665

667 Table 1. Accession, locality and assembly information for samples included in this study.

Species	Accession	Locality (patch size)	Assembly
<i>B. bracteosa</i>	NGW8873	Dargo Rd, Dargo VIC	3
<i>B. grayi</i>	CBG7800854	Cultivated: RBGV (CANB741502) Origin: Cotter pumping station, Murrumbidgee River ACT	3
<i>B. riparia</i>	RBGV-070563	Cultivated: RBGV-030506. Origin: Big River 7.5 km SW of Jamieson	3
<i>B. vombata</i>	BV31	Wombat State Forest, VIC, site 1 (9x11 m)	NA, low reads
<i>B. vombata</i>	BV32	Wombat State Forest, VIC, site 1 (9x11 m)	1,2
<i>B. vombata</i>	BV33	Wombat State Forest, VIC, site 1 (9x11 m)	NA, low reads
<i>B. vombata</i>	BV34	Wombat State Forest, VIC, site 1 (9x11 m)	NA, low reads
<i>B. vombata</i>	BV35*	Wombat State Forest, VIC, site 1 (9x11 m)	1,2,3
<i>B. vombata</i>	BV36*	Wombat State Forest, VIC, site 2 (2x4 m)	1
<i>B. vombata</i>	BV37	Wombat State Forest, VIC, site 2 (2x4 m)	1,2,3
<i>B. vombata</i>	BV38	Wombat State Forest, VIC, site 2 (2x4 m)	1
<i>B. vombata</i>	BV39	Wombat State Forest, VIC, site 2 (2x4 m)	1,2
<i>B. vombata</i>	BV40	Wombat State Forest, VIC, site 2 (2x4 m)	1,2
<i>B. vombata</i>	BV41	Wombat State Forest, VIC, site 3 (5x11 m)	1,2
<i>B. vombata</i>	BV42	Wombat State Forest, VIC, site 3 (5x11 m)	1,2
<i>B. vombata</i>	BV43	Wombat State Forest, VIC, site 3 (5x11 m)	1
<i>B. vombata</i>	BV44	Wombat State Forest, VIC, site 3 (5x11 m)	1,2,3
<i>B. vombata</i>	BV45	Wombat State Forest, VIC, site 3 (5x11 m)	1
<i>B. vombata</i>	BV51	Wombat State Forest, VIC, site 4 - patch 1 (9x8 m)	1
<i>B. vombata</i>	BV52*	Wombat State Forest, VIC, site 4 - patch 1 (9x8 m)	1
<i>B. vombata</i>	BV53	Wombat State Forest, VIC, site 4 - patch 1 (9x8 m)	1
<i>B. vombata</i>	BV54	Wombat State Forest, VIC, site 4 - patch 1 (9x8 m)	1,2,3
<i>B. vombata</i>	BV55	Wombat State Forest, VIC, site 4 - patch 1 (9x8 m)	1
<i>B. vombata</i>	BV46	Wombat State Forest, VIC, site 4 - patch 2 (3x4 m)	1
<i>B. vombata</i>	BV47	Wombat State Forest, VIC, site 4 - patch 2 (3x4 m)	1
<i>B. vombata</i>	BV48	Wombat State Forest, VIC, site 4 - patch 2 (3x4 m)	1
<i>B. vombata</i>	BV49	Wombat State Forest, VIC, site 4 - patch 2 (3x4 m)	1
<i>B. vombata</i>	BV50	Wombat State Forest, VIC, site 4 - patch 2 (3x4 m)	1,3
<i>B. vombata</i>	BV56	Wombat State Forest, VIC, site 4 - patch 3 (5x13 m)	1
<i>B. vombata</i>	BV57	Wombat State Forest, VIC, site 4 - patch 3 (5x13 m)	1
<i>B. vombata</i>	BV58	Wombat State Forest, VIC, site 4 - patch 3 (5x13 m)	1,2,3
<i>B. vombata</i>	BV59*	Wombat State Forest, VIC, site 4 - patch 3 (5x13 m)	1
<i>B. vombata</i>	BV60	Wombat State Forest, VIC, site 4 - patch 3 (5x13 m)	1
<i>B. vombata</i>	BV61*	Wombat State Forest, VIC, site 4 - patch 4 (17x6 m)	1
<i>B. vombata</i>	BV62*	Wombat State Forest, VIC, site 4 - patch 4 (17x6 m)	1,2
<i>B. vombata</i>	BV63	Wombat State Forest, VIC, site 4 - patch 4 (17x6 m)	1,2
<i>B. vombata</i>	BV64	Wombat State Forest, VIC, site 4 - patch 4 (17x6 m)	1
<i>B. vombata</i>	BV65	Wombat State Forest, VIC, site 4 - patch 4 (17x6 m)	1,2,3
<i>B. vombata</i>	BV66*	Wombat State Forest, VIC, site 4 - patch 5 (14x6 m)	1,3
<i>B. vombata</i>	BV67	Wombat State Forest, VIC, site 4 - patch 5 (14x6 m)	1
<i>B. vombata</i>	BV68*	Wombat State Forest, VIC, site 4 - patch 5 (14x6 m)	1
<i>B. vombata</i>	BV69	Wombat State Forest, VIC, site 4 - patch 5 (14x6 m)	1
<i>B. vombata</i>	BV70	Wombat State Forest, VIC, site 4 - patch 5 (14x6 m)	1
<i>B. walkeri</i>	RBGV-070595*	Cultivated: RBGV. Origin: unknown. Natural distribution NW Victoria, western New South Wales, SW South Australia, SW Western Australia	3

668 Assembly numbers represent 1) clonal identification, 2) chloroplast haplotype & 3) phylogenetic assembly.

669 Asterisk reflects accession with technical replicate pair.

670 Table 2: Pairwise population differentiation (G_{ST} : lower left, Jost's D : upper right) among five
671 identified clones of *Bossiaea vombata*. Calculations are based on 720 ddRADseq SNPs.

	Clone 1	Clone 2	Clone 3	Clone 4	Clone 5
Clone 1	-	0.091	0.113	0.113	0.057
Clone 2	0.340	-	0.139	0.124	0.091
Clone 3	0.423	0.508	-	0.140	0.112
Clone 4	0.453	0.499	0.558	-	0.050
Clone 5	0.223	0.350	0.423	0.253	-

672

673

674 **Figure Captions**

675 Fig 1. Distribution map showing (a) the location of the Wombat State Forest, Victoria, Australia
676 and (b,c) indicative locations of *Bossiaea vombata* individuals collected during this study.
677 Maps were generated using Google's 'terrain' map data in R via the (a) 'maps' and (b)
678 'ggmap' packages.

679 Fig 2. Genetic divergence threshold identification (dashed line) among five *Bossiaea vombata*
680 clones in the Wombat State Forest, Victoria, Australia. Analyses are based on 720 SNPs.
681 The x-axis represents genetic distance. The y-axis represents; (a) pairwise comparisons of
682 genetic distance for each SNP and (b) a summary of pairwise comparison occurrences
683 (density) for a given genetic distance. We find (a, b) a distinct difference between genetic
684 diversity representing library preparation, sequencing error and somatic mutations (left
685 of dashed line) and that representing inter-clonal variation (right of dashed line) and (c)
686 the presence of five multi-locus lineages when applying this threshold to an ultrametric
687 topology based on genetic distances. Sites 1-3 were monoclonal (clone 1-3, respectively).
688 Site 4 comprised clones 4 and 5.

689 Fig 3. DAPC plot of *Bossiaea vombata* individuals based on 720 SNPs obtained via ddRADseq with
690 no input regarding a priori site information. The x-axis represents discriminant function
691 one, which accounts for 52% of the overall variability in our dataset. The y-axis represents
692 discriminant function two and accounts for 22% of the overall genetic variability. (a) Here,
693 we see five distinct clusters representing independent clonal lineages. (b) Graphical
694 representation showing membership assignment probabilities (y-axis) for 37 individuals
695 (x-axis) as calculated via DAPC analysis. Sites 1-3 were monoclonal (clone 1-3,
696 respectively). Site 4 comprised clones 4 and 5.

697 Fig 4. Unrooted phylogram of five leafless *Bossiaea* species. **Maximum Likelihood analysis**
698 **was** based on 2,893 SNPS obtained via ddRADseq. **Bootstrap support values above 85 are**
699 **shown at each node.** Distributions based on Thompson (2012) and voucher information
700 from The Australasian Virtual Herbarium (<https://avh.ala.org.au>).

